



## Ocean acidification impact on the grooved carpet shell clam (*Ruditapes decussatus*)

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### ABSTRACT

The grooved carpet shell clam (*Ruditapes decussatus*) is one of the most economically important mollusks inhabiting Mediterranean lagoons and sandy beaches both from fisheries and aquaculture. The present study aims to study the impact of different levels of acidification on this calcifying organism. Juvenile clams (avg. Shell Length, SL= 23.22 ± 0.84 mm) were incubated in CO<sub>2</sub> enriched seawater at four different CO<sub>2</sub> concentrations [420 ppm (ambient control), 550 ppm, 750 ppm and 1050 ppm] representing projected atmospheric CO<sub>2</sub> concentration scenarios for the year 2100 by IPCC. The studied biological parameters showed slight decrease with increasing pCO<sub>2</sub>. However, differences were not significant. Standard length decreased as pCO<sub>2</sub> concentration increased, with a maximum average decrease of (-0.12) recorded at 750 ppm as compared to the control group. Regarding total weight, the decrease was highest (-0.10) in both 550 and 1050 ppm. Moreover, clams kept at 550 ppm showed the lowest condition index (11.40 ± 1.49) and highest mortality rate of 8%. The study of physiological response showed increase in metabolic rate and ammonia excretion in both 550 ppm and the control 420 ppm groups. Algal feed clearance rate decreased with increasing acidification with highest value in the control (420 ppm) group and lowest average value of 3.34 l/h<sup>-1</sup> in the extremely high pCO<sub>2</sub> (1050 ppm) group. By the end of century, ocean acidification may exert additional stress on the health of *R. decussatus* and its economic value.

### INTRODUCTION

Ocean acidification is the reduction of pH in the seawater due to the increased levels of atmospheric CO<sub>2</sub>. At a quickened rate, the CO<sub>2</sub> deposition into the atmosphere has been increased since the industrial revolution (beginning in the 1750s). The increases in atmospheric carbon dioxide partial pressure (pCO<sub>2</sub>), and different greenhouse gases have been given a share in global warming. The Current average atmospheric pCO<sub>2</sub> of approximately 400 ppm is higher than the highest pCO<sub>2</sub> over the past 2.1 million years (approximately 300 ppm; Hönisch *et al.* 2009). According to the Intergovernmental Panel on Climate Change's (IPCC) B1 scenario

(stabilized population growth by 2050 and a more service- and information-driven economy), proposed average global CO<sub>2</sub> emissions will attain 650 ppm by 2100 (Caldeira and Wickett 2005). Following Henry's law, the increase in atmospheric pCO<sub>2</sub> will result in an increase in its concentration in an adjacent water body as a result of physical dissolution through the surface microlayer. About a quarter of the CO<sub>2</sub> produced daily by human activities is being absorbed by the oceans, causing ocean acidification which results from changing the seawater carbonate system chemical equilibrium. The dissolution of CO<sub>2</sub> in oceans causes increase in acidity (decrease in pH) and reduction in the obtainable carbonate ions (CO<sub>3</sub><sup>2-</sup>), and consequently become increasingly shallower under saturation horizon of CaCO<sub>3</sub> (Feely *et al.*, 2004, 2012), which is projected to become more persistent and widespread by 2050 and beyond, (Gruber *et al.*, 2012). If population growth occurs at a higher rate accompanied by slow economic development and limited technological changes (IPCC's A2 scenario), emissions will likely reach 970 ppm by 2100 (Caldeira and Wickett 2005). Ocean acidification is expected to affect ecosystems at an accelerating pace over the next century (Caldeira and Wickett 2003; IPCC 2007).

CaCO<sub>3</sub> is required by marine organisms as the main building blocks of their skeletons, shells and other calcareous frames. In 250 years, Ocean acidity has increased by 30%, which is equivalent to a decreasing of surface seawater pH of 0.1 units (Orr *et al.*, 2005), therefore, the proposed scenarios by 2100 could be tripled.

pH has a vital role in all marine organisms and the change in internal pH can affect an organism's health or even lead to death. Marine organisms have to preserve their internal pH prorated to that of the surrounding seawater. Some have complicated systems that can regulate internal pH. The most heavily influenced species by their surrounding environment are the ones without these systems and can be rapidly menaced by changes in acidity. Although the research concerned with the impact of ocean acidification on marine environment is growing very fast, still poorly understood and the picture is not fully clear. However, few studies identified some marine organisms that can be threatened by changing the environment (Gazeau *et al.*, 2007; Beniash *et al.*, 2010; Michaelidis *et al.*, 2005; Thomsen and Melzner, 2010, Ibrahim *et al.*, 2014, Khairy *et al.*, 2014).

The carpet shell clam *Ruditapes decussatus* is rated as one of the most popular bivalves with a high economic value in many countries (Chessa *et al.* 2005; Prado-Alvarez *et al.*, 2009; Lucrezia *et al.* 2011). It is intended as an auspicious candidate for the emerging bivalve aquaculture progression in Egypt (El-Wazzan *et al.* 2012; Abbas *et al.* 2018). The present study aims to study the impact of different levels of acidification on this calcifying organism.

## MATERIALS AND METHODS

### Study organism collection

The grooved carpet shell clam *Ruditapes decussatus* was collected in November 2017 from the Lake Timsah in Ismailia, Egypt, which is located on the Suez Canal at 30°34'N and 32°18'E. Clams of experiments were transported and placed in the receiving tank to recover from the transportation stress and acclimate to lab conditions in wet laboratories at the National Institute of Oceanography and Fisheries – Alexandria, Egypt. The mean length at the start of the study was 23.22 mm (±0.84 SD).

### Experimental design

The experimental design was composed of four treatments representing different  $\text{PCO}_2$  (420 ppm, 550 ppm, 750 ppm, 1050 ppm). After two days of acclimation, clams were housed in 3 L of  $1\mu\text{m}$  filtered seawater in 4-liter tanks. The experiment was running in two sets A and B with exactly the same condition. Each set was composed of 12 experimental tanks (3 replicate tanks per each  $\text{PCO}_2$  treatment). Each tank had an air pump for providing oxygen for healthy circulation system. Set A was used to measure metabolic rate, clearance rate and ammonia excretion, while set B had been set up for mortality recording, shell length, total weight and condition index measurements.

Clams were fed with *Isochrysis galbana* at concentration of 100,000 cell/ml once daily. Clams were incubated for a period of 36 days at the four  $\text{pCO}_2$  concentrations mentioned above keeping alkalinity constant.

### **Seawater chemistry manipulation**

Seawater was adjusted to the four  $\text{CO}_2$  concentrations {420 ppm (ambient control), 550 ppm, 750 ppm and 1050 ppm} and the corresponding pH's by mixing with  $\text{CO}_2$  saturated sea water while keeping alkalinity constant by using the seacarb package within R program. This  $\text{CO}_2$  saturated seawater was acidified through bubbling of  $1\mu$  filtered sea water with pure  $\text{CO}_2$  gas till saturation. This design for acidification experiment manipulation had declared the best way to manipulate the  $\text{pCO}_2$  at constant alkalinity.

Seawater sampled for the initial pH was measured by pH meter (Jenway 3505). The pH electrode was calibrated with TRIS buffer on a total (T) scale following Dickson *et al.* (2007). Total alkalinity (TA) was measured following Sarazin *et al.* (1999). Certified reference materials (CRM batch 115) were used to calibrate and establish correction factors for TA measurements that were obtained from Professor Andrew Dickson at the Marine Physics Laboratory of the Scripps Institute of Oceanography, University of California, San Diego. Seawater temperature, salinity, silicate and phosphate were measured spectrophotometry according to Grasshoff *et al.* (1983) and used as an input for  $\text{CO}_2$  sys calculations.

Seawater carbonate chemistry parameters in the beginning of the experiment and throughout it were determined via calculations by  $\text{CO}_2\text{SYS}$  using the two measured  $\text{CO}_2$  parameters,  $\text{pH}_{\text{sw}}$  and  $\text{TA}_{\text{sw}}$ . Regarding the carbonate system, dissociation constants  $K_1$  and  $K_2$  (Mehrbach *et al.*, 1973, refitted by Dickson and Millero, 1987) were used. The acidity constant of the  $\text{H}_2\text{SO}_4$  in seawater was calculated using the constants of Dickson (1990).

Water chemistry in each system was replaced daily, monitored and adjusted every 2 hours using carbon dioxide saturated Seawater.

### **Biometrics measurements**

#### **Shell length and total weight:**

Shell lengths of 360 individuals (30 individuals per replicate jar x 3 replicates for each of 4 treatments:  $30 \times 3 \times 4 = 360$ ) were measured initially and at the end of the experiment using a Vernier caliper to 0.1 mm. In clams, the length corresponds to the anterior/posterior axis and is measured perpendicularly from the height line, matching the dorsal/ventral axis. Weights of the same group were also measured using four decimals  $\pm 0.0002$  ordinary laboratory balance.

#### **Condition index (C.I):**

A day before setting up the experiment, condition index (C.I) was determined for 50 individuals in pre-weighted pans. The measurements of C.I. were repeated at

the end of the experiment for 30 individuals per treatment and its triplicate, the clams were dissected after their shell length and total weight had been recorded, they were opened by cutting the adductor muscle, the shell and viscera were entirely separated in their own weighed pans, both were dried in separate pans at 90°C for 24 h, then they were weighed again. C.I was calculated according to (Walne 1976) as follows,  $C.I = [(dry\ flesh\ weight / dry\ shell\ weight) \times 100]$ .

#### **Mortality:**

Mortality was daily recorded by counting dead and gapping (dying) clams after which they were removed from the jars and their shell lengths measured and recorded by date.

#### **Physiological measurements**

##### **Metabolic rate:**

Metabolic rate was calculated by measuring oxygen consumption rate (OCR) by incubating one clam per jar from each treatment in 250 mL oxygen bottle for 3h. Clams were not fed for a period of 6h before the start of the incubations. Oxygen consumption rates were estimated as the rate of oxygen concentrations decrease over the 3h incubations as measured according to the modified Winkler method for initial and after 3h incubation time. OCR was calculated using the following equation (Cerezo Valverde *et al.* 2006):

$$OCR = (DO_0 - DO_t) V / (DW T)$$

The initial and final concentrations of dissolved oxygen (DO) are denoted by subscripts 0 and t time respectively, V is the volume of respiration chamber (l), DW is the dry weight of *R. decussates*, and T is the time between the initial and final measurements (h).

##### **Clearance rate:**

Clearance rate was determined through the volume of water cleared of particles per unit time, and it can be determined indirectly by monitoring the decline in algal cells in a "closed system". After allowing 5 minutes for the algal cells to mix thoroughly. A sample of 10 ml was taken from the center of each jar with a 10-mL pipette. Samples were collected after 2 hours. The cell concentrations were counted immediately using an electronic particle counter Coulter Counter® Model ZM or D, fitted with a 140 µm aperture tube. Cell concentrations were the mean of 3 to 4 counts. Clearance rate was calculated using the equation (Coughlan, 1969):

$$CR = V (\ln C_1 - \ln C_2) / t$$

Where: CR is the clearance rate, V is the volume of water used,  $C_1$  and  $C_2$  are the cell concentrations between two sampling times, and t is the time increment.

##### **Ammonia excretion rate:**

The experiment was conducted by placing 30 adult clams (~ 6 g dry weight) in 3 L of seawater and monitoring the ammonia concentration initially and 24 hours after feeding with *I. galbana*. This analysis was performed according to methods described by Grasshoff *et al.* (1983).

##### **Statistical analysis:**

All statistical analyses were performed using SPSS (version 22). Paired T- test has been performed to declare the differences between before and after the experiment in both shell length and total weight of the studied clams. Moreover, a one-way ANOVA was performed with treatment as a single factor to determine differences among treatments (different acidification levels) and effect of time post exposure on different indices of clam physiological responses.

## **RESULTS**

Mean size of clams used in the present study was  $SL = 23.22 \text{ mm} \pm 0.84 \text{ SD}$ . pH of the ambient control represented by  $pCO_2$  of 420 was about  $8.12 \pm 0.06$ , and for those representing ocean acidification conditions of 550 ppm, 750 ppm and 1050 ppm were  $7.98 \pm 0.05$ ,  $7.84 \pm 0.04$  and  $7.62 \pm 0.06$  respectively.

#### The impact of ocean acidification on shell length:

Comparing different  $pCO_2$  treatments, *Ruditapes decussatus* showed no significant changes in their shell length between SL before the start and at the end of the experiment ( $P \geq 0.05$ ; Paired t- test in each treatment) with no significant differences in SL among all treatments including the control after exposure to different acidification levels (ANOVA shown in Table 1).

#### The impact of ocean acidification on total weight:

Total weight had insignificant decrease amongst different treatments ( $P \geq 0.05$ ; Table 1) after 36 days of acidification experiment at different  $pCO_2$  concentrations. Paired t-test showed that there were significant differences between total before and after the experiment for both treatments 550 and 1050 ppm ( $P = 0.024$  and  $0.002$ , respectively).

Table 1: One-way ANOVA testing the biological and physiological parameters in *Ruditapes decussatus* clams exposed to different acidification conditions ( $pCO_2 = 420, 550, 750$  and  $1050$  ppm).

| Source                              | df | F      | Sig.  |
|-------------------------------------|----|--------|-------|
| Shell length                        | 3  | 1.226  | 0.300 |
| Total weight                        | 3  | 3.809  | 0.010 |
| Condition Index                     | 4  | 23.299 | 0.000 |
| Metabolic Rate (concentration)      | 3  | 0.445  | 0.723 |
| Metabolic Rate intervals (TIMES)    | 8  | 7.122  | 0.000 |
| Clearance Rate (concentration)      | 3  | 0.681  | 0.572 |
| Clearance Rate intervals (TIMES)    | 7  | 3.818  | 0.006 |
| Ammonia Excretion (concentration)   | 3  | 1.349  | 0.326 |
| Ammonia Excretion intervals (TIMES) | 3  | 4.1    | 0.051 |

#### The impact of ocean acidification on condition index:

Condition index values decreased in all treatments compared to the initial values of the stock (Fig. 1).

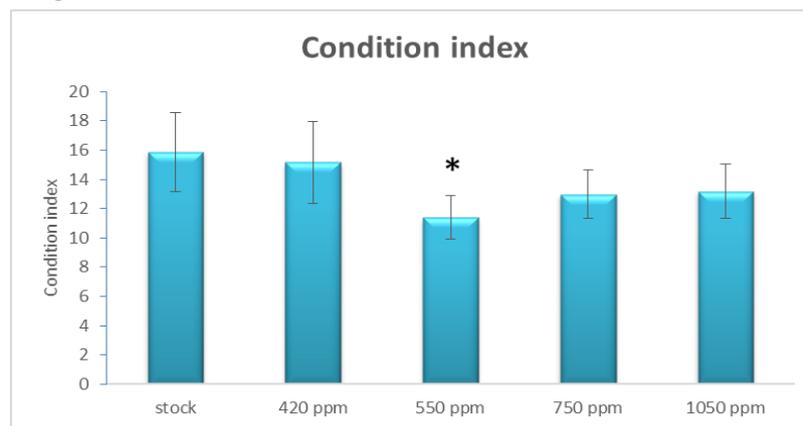


Fig. 1: Mean condition index (gDW  $\pm$ SD) of *Ruditapes decussatus* exposed to different ocean acidification conditions represented by different  $pCO_2$  (420 ppm (ambient control), 550 ppm, 750 ppm and 1050 ppm) for 36 days experiment. Asterisk (\*) represent significant difference.

Result showed that the lowest condition index of  $11.40 \pm 1.49$  was observed in the 550 ppm treatment, while the ambient control group showed C.I value of  $15.17 \pm 2.80$ . Condition index had significant differences among different treatments ( $P < 0.05$ ; Table 1).

### The impact of the ocean acidification on clam mortality:

No significant effect of ocean acidification was observed on clam mortality. Mortality of *Ruditapes decussatus* incubated clams at 550 ppm had high mortality of 8% compared to the ambient group (3%) which is not significant in both cases. No other high mortalities occurred in any of the treatments (Fig. 2).

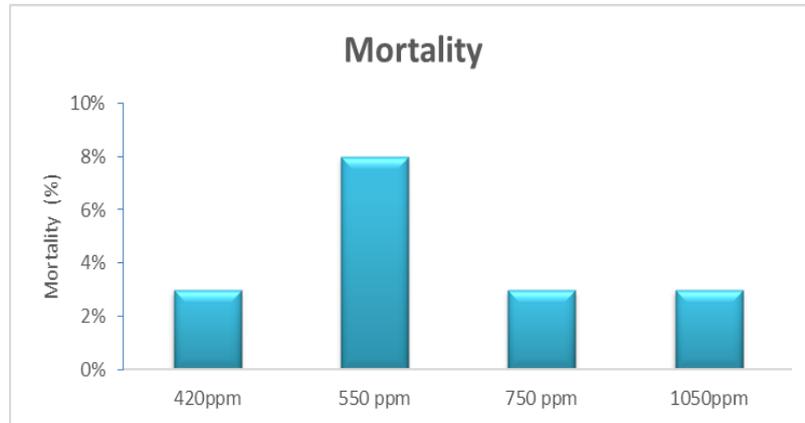


Fig. 2. Mortality (%) of *Ruditapes decussatus* exposed to different ocean acidification conditions represented by different pCO<sub>2</sub> (420 ppm (ambient control), 550 ppm, 750 ppm and 1050 ppm) for 36 days' experiment.

### The impact of the ocean acidification on metabolic rate:

In calculating oxygen consumption rate, a high metabolic rate was observed at the beginning of the experiment in all treatments (Fig. 3). Metabolic rates fluctuated throughout the experimental duration but in general was lower than initial value. However, It did not show significant difference among different groups (Table 1). Briefly; following the metabolic rate over the experimental time intervals, metabolic rates of clams were lower than initial values after 5 days and 9, 31 and 36 days as compared to initial values (Fig. 3). Time post exposure to different acidification levels had significant impact on metabolic rate in clams (P=0.000).

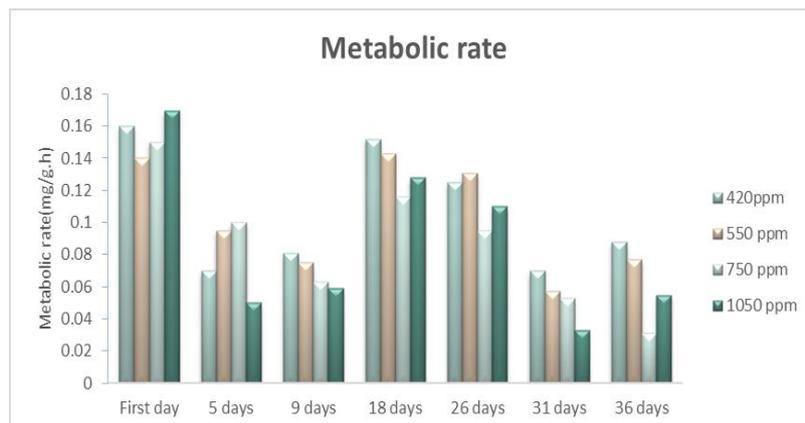


Fig. 3. Metabolic rate of clam (*Ruditapes decussatus*) batch (A) under different acidification conditions (pCO<sub>2</sub> = 420, 550, 750 and 1050 ppm).

### The impact of the ocean acidification on clearance rate:

The algae clearance rates by *Ruditapes decussatus*, at different acidification concentration, showed no significant differences among treatments (Table 1). Increasing the time of exposure to acidification showed significant differences among

time intervals ( $P= 0.006$ ). Clams incubated at  $p\text{CO}_2$  (1050 ppm) showed the lowest clearance, while the highest clearance rate was recorded by the ambient group  $p\text{CO}_2$  (420 ppm; Fig. 4).



Fig. 4: Clearance rate of algae by *Ruditapes decussatus* under different acidification conditions ( $p\text{CO}_2 = 420, 550, 750$  and  $1050$  ppm).

#### The impact of the ocean acidification on Ammonia excretion:

Ammonia excretion determination showed that the highest ammonia concentration was recorded in the groups of 420 ppm ambient control and 550 ppm, while the lowest ammonia concentration was recorded in the 1050 ppm group (Fig. 5). However, ANOVA test recorded no significant differences between different groups (Table 1). Moreover, ANOVA cleared that there was a significant effect of time of exposure to acidification ( $P= 0.05$ ; Table 1).

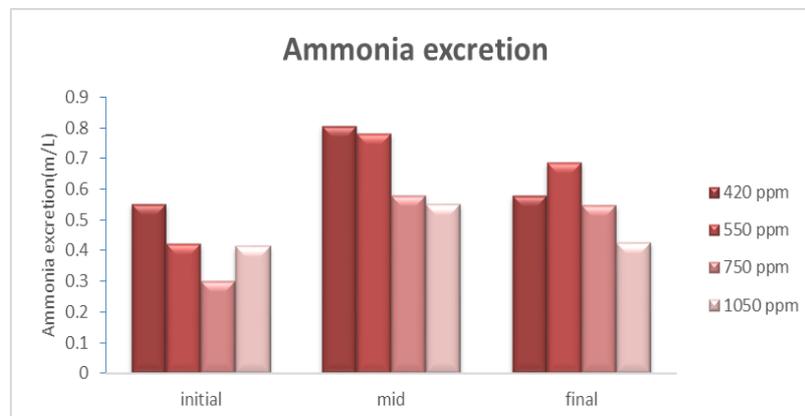


Fig. 5: Ammonia excretion of *Ruditapes decussatus* under different acidification conditions ( $p\text{CO}_2 = 420, 550, 750$  and  $1050$  ppm).

## DISCUSSION

The present study is focused on the effect of ocean acidification on the Mediterranean clam *Ruditapes decussatus* which is rated as one of the most commercially important and highly economic bivalve in many countries including Egypt (FAO 2018; Abbas *et al.* 2018). This is due to increasing consumer demand and significance for potential aquaculture of bivalve in Egypt.

As the ascent within the ocean acidification field escalates by time, the impacts of lowered seawater pH are hugely affecting the energy distribution requirements of the clams, as more energy is needed for the conservation processes of the

physiological balance (Wood *et al.* 2008, Thomsen and Melzner 2010, Stump *et al.* 2011, Pan *et al.* 2015).

Studies that examined bivalve's responses to ocean acidification showed that adult bivalves are mostly less sensitive to changes in oceanic pCO<sub>2</sub> than the larval stages, but they still display changes in growth, calcification, and physiological responses (Dupont *et al.*, 2010). Bivalve shell deposition is negatively impacted by elevated pCO<sub>2</sub> (Gazeau *et al.*, 2007), and some studies have illustrated decreased shell growth in low pH-exposed juveniles and adults (Beniash *et al.*, 2010; Michaelidis *et al.*, 2005; Thomsen and Melzner, 2010).

#### **The impact of the ocean acidification on biometric measurements:**

Growth of marine organisms is a physiological process which becomes significantly slower under long-term exposure of increased CO<sub>2</sub> levels (Michaelidis *et al.*, 2005).

Some studies have shown a link between pH conditions and bivalve growth rates. Researches are pointing to the inhibiting effect of ocean acidification on the growth of different species of bivalve (Michaelidis *et al.*, 2005, Berge *et al.*, 2006). For example, (Shirayama and Thornton, 2005) found that even very average increases in CO<sub>2</sub> of 200 ppm above present levels lead to reduction in the growth rate and the survival rate of echinoderms and gastropods, indicating that long-term of exposure to CO<sub>2</sub> changes can affect the growth of calcifying organisms.

Bressan *et al.* (2014) found that under acidified conditions, the *Chamelea gallina* did not grow, but their live weight decreased greatly and a slight reduction in shell length was even observed.

In the present study, the impact of increased CO<sub>2</sub> on growth is represented in terms of shell length and weight. SL changes showed a slight insignificant reduction in shell size. Similarly, experiments on juvenile *Veneridae* (Venus clam) have not shown any significant difference in growth which is measured as size, weight and net calcification (Range *et al.*, 2011, Talmage and Gobler, 2011). Moreover, the studies on *Mytilus edulis* (Melzner *et al.* 2011) showed no significant CO<sub>2</sub> effect to shell mass and length. The studies of the impact acidification on shell length and shell weight of *Mytilus galloprovincialis* (Gazeau *et al.*, 2014) did not appear to be highly sensitive to ocean acidification. Similar recorded results on the clam *Mercenaria mercenaria* were explained as a result of shell erosion (Ringwood and Keppler, 2002).

Regarding the recorded total weight during the current experiment, results showed no positive growth was recorded in terms of total live weight in the clams, with no significant differences being detected between the control and treated animals. Other studies described that growth might be significantly reduced under medium and long-term exposure to elevated pCO<sub>2</sub> levels in mussel *Mytilus chilensis* (Navarro *et al.*, 2013; Duarte *et al.*, 2014). Moreover, both shell and soft tissue growth of the oyster *C. virginica* were reduced when exposed to high pCO<sub>2</sub> levels (pH 7.5; Beniash *et al.*, 2010).

The recorded results in the current study of the C.I. values decreased throughout all treatments compared to the ambient control group. Result showed the lowest condition index ( $11.40 \pm 1.49$ ) at 550 ppm, compared to the control group ( $15.17 \pm 2.80$ ) that was approximately unchanged from the stock. The high condition values are good indicators for nutrient reserves accumulation for total weight growth in bivalves (Filgueira *et al.* 2013). The lowest recorded condition index (C.I) occurred at 550ppm CO<sub>2</sub> where it is consistent with the highest mortality observation. This may explain the observed lower energy allocation to growth under pCO<sub>2</sub> conditions (550 and 1050ppm).

Mortality (%) of clam *Ruditapes decussatus* in the present study showed a maximum mortality of 8% recorded at 550 ppm and delayed mortality at higher pCO<sub>2</sub> (750 and 1050 ppm). This could be explained according to the hypothesis of Gazeau *et al.* (2014) by which exerting metabolic depression CO<sub>2</sub> can alleviate the level of stress and delay mortality through more efficient exploitation of energy reserved and passive tolerance as an adaptation strategy. This is achieved through more efficient imposition of energy reserves and passive tolerance. Therefore, the results of the current study showed low metabolic rate and low mortality rate recorded at both treatments (750 and 1050 ppm), whereas group 550 ppm showed higher mortality (8%) as they had high metabolic rate relative to the ambient control group.

#### **The impact of the ocean acidification on physiological responses:**

Several studies have described metabolic depression in different species of marine organisms at elevated pCO<sub>2</sub> concentrations as caused by the low capacity to compensate for disturbances in extracellular ion and acid–base status (Michaelidis *et al.*, 2005; Pörtner, 2008). This is in agreement with the results of the present study, where the oxygen uptake was depressed at 750 and 1050 ppm CO<sub>2</sub> as compared to the control seawater. A depression in the metabolism caused by the un neutralized acid–base status and produce “trade-off” in energy budget in many species. This adaptation strategy is utilized to match Adenosine triphosphate supply and Adenosine triphosphate demand and thus extend their survival (Melatunan *et al.* 2013). On the other hand, the clams under 550ppm CO<sub>2</sub> showed high metabolic rate associated with to the highest mortality as explained previously. However, this was referred to as an increase in food absorption efficiency (Fernández-Reiriz *et al.*, 2012). Fernández-Reiriz *et al.* (2011) reported similar results for the specific rate of oxygen consumption by juveniles of the clam *R. decussatus*, with lower values in individuals exposed to high levels of pCO<sub>2</sub> with ( $\Delta$  pH -0.7).

Considering the impact of sea water acidification on clams feeding, the results showed a pronounced decrease in the clearance rate at the highest pCO<sub>2</sub> level (1050 ppm) corresponding to pH of (7.62). The negative effect of increasing pCO<sub>2</sub> on the feeding rate of *R. decussatus* observed in the present study had also been recorded for the same species (Bamber 1987; Fernandez-Reiriz *et al.* 2011). Fernandez-Reiriz *et al.* (2011) observed a reduction in the feeding rate in the clam *Ruditapes decussatus* at the highest experimental pCO<sub>2</sub> equivalent to pH 7.48. Similarly, according to Bamber (1987), the feeding activity of the clam *R. decussatus* is inhibited at pH < 7.0, and both, tissue and shell growths were significantly reduced. Bamber (1990) also described a negative effect of seawater acidification on the feeding activity and growth of the bivalves *Ostrea edulis*. Results of the present study may be explained by indicating possible deficiencies in the functioning of the digestive systems under conditions of seawater acidification which is synergized with metabolic decline, where the oxygen uptake decreased at elevated pCO<sub>2</sub> levels with high pCO<sub>2</sub> levels in the seawater (Navarro *et al.*, 2013).

Several studies reported low ammonia excretion (AE) of marine bivalves with elevated pCO<sub>2</sub> levels (Bayne and Newell, 1983; Velasco and Navarro, 2005). Considering the impact of seawater acidification on clam ammonia Excretion (AE), Wang *et al.* (2015) observed low values for AE (20–45%) in the mussel *Mytilus coruscus* with no significant differences between three different pH levels (8.1, 7.7, and 7.3). Navarro *et al.* (2013) reported that elevated pCO<sub>2</sub> levels (pH 7.57) significantly reduced the AE of *Mytilus chilensis*, indicating possible deficiencies in the functioning of the digestive system under conditions of seawater acidification. In the present study, Ammonia excretion was affected by acidification, and a decrease

was observed in clams exposed to the highest pCO<sub>2</sub> (1050 ppm; pH 7.62). These results are in consistence with the recorded low clearance rate and low metabolic rate in sequence.

## CONCLUSION

The present study suggests that ongoing ocean acidification may not pose great threats to the existence of Mediterranean clam *Ruditapes decussatus* as many scenarios expected. The poor, clearance rate (feeding rate), growth and condition index upon exposure to ocean acidification may lead to poor aquaculture potential and health of *R. decussatus* unless adaptation mechanisms may develop with continuous gradual exposure to increasing ocean acidification. Risk assessment is needed for the Mediterranean bivalve aquaculture in the current century. More research is needed on consequent generations to investigate the possibility of the organisms to create adaptation strategies convenient to future climate scenarios at both organismal and genetic levels.

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## REFERENCES

- Abbas, A.S.A.; El-Wazzan, E.; Khafage, A.R.; El-Sayed, A.M. and Abdel Razek, F.A. (2018). Influence of different microalgal diets on gonadal development of the carpet shell clam *Tapes decussatus* broodstock. *Aquaculture International*. 26(5):1297–1309. DOI: <https://doi.org/10.1007/s10499-018-0284-9>.
- Bamber, R. (1987). The effects of acidic sea water on young carpet-shell clams *Venerupis decussata* (L.) (Mollusca: Veneracea). *Journal of Experimental Marine Biology and Ecology*. 108(3): 241-260.
- Bamber, N.R. (1990). The effects of acidic sea water on three species of lamellibranch molluscs. *Journal of Experimental Marine Biology and Ecology*. 143(3): 181-191. [https://doi.org/10.1016/0022-0981\(90\)90069-O](https://doi.org/10.1016/0022-0981(90)90069-O).
- Bayne, B.L. and Newell, R.C. (1983). Physiological energetics of marine molluscs. In *The Mollusca. Physiology*. 4(1): 407–515.
- Beniash, E.; Lieb, N.I.; Kurochkin, I. and Sokolova, I. (2010). Elevated level of carbon dioxide affects metabolism and shell formation in oysters *Crassostrea virginica* (Gmelin). *Marine Ecology Progress Series*, 419: 95–108. doi:10.3354/meps08841
- Berge, J.A.; Bjerkeng, B.; Pettersen, O. ; Schaanning, M.T., and Oxnevad, S. (2006). Effects of increased seawater concentrations of CO<sub>2</sub> on growth of the bivalve *Mytilus edulis* L. *Chemosphere*. 62(4): 681–687. <https://doi.org/10.1016/j.chemosphere.2005.04.111>
- Bressan, M.; Chinellato, A.; Munari, M.; Matozzo, V.; Mancini, A. and Marceta, T. (2014). Does seawater acidification affect survival, growth and shell integrity in bivalve juveniles? *Marine Environmental Research* 99: 136-148. <https://doi.org/10.1594/PANGAEA.836888>.

- Caldeira, K.; and Wickett, M.E. (2003). Anthropogenic carbon and ocean pH. *Nature*. 425: 365. doi: 10.1038/425365a.
- Catarino, A.I.; De Ridder, C.; Gonzalez, M.; Gallardo, P. and Dubois, P. (2012). Sea urchin *Arbacia dufresnei* (Blainville 1825) larvae response to ocean acidification. *Polar Biology*. 35(3): 455-461. doi: 10.1007/s00300-011-1074-2
- Caldeira, K. and Wickett, M.E. (2005). Ocean model predictions of chemistry changes from carbon dioxide emissions to the atmosphere and ocean. *Journal of Geophysical Research*, 110(C9): 1-12. doi:10.1029/2004JC002671.
- Cerezo, V. J.; Martínez L. F.J.; and García G. B. (2006). Oxygen consumption and ventilatory frequency responses to gradual hypoxia in common dentex (*Dentex dentex*): Basis for suitable oxygen level estimations. *Aquaculture* 256(1-4): 542–551.
- Chessa, L.A.; Paesanti, F.; Pais, A.; Scardi, M.; Serra, S. and Vitale, L. (2005). Perspective for development of low impact aquaculture in western Mediterranean lagoon: the case of the carpet clam *Tapes decussatus*. *Aquaculture International* 13: 147-155. doi.org/10.1007/s10499-004-9022-6
- Coughlan, J. (1969). The estimation of filtering rate from the clearing of suspension. *Marine Biology*. 2: 356-358. 10.1007/BF00355716.
- Dickson, A.G. and Millero, F.J. (1987). A comparison of the equilibrium constants for the dissociation of carbonic acid in seawater media. *Deep Sea Research Part A, Oceanographic Research Papers*, 34(10): 1733-1743. Doi:10.1016/0198-0149(87)90021-5
- Dickson, A.G. (1990b). Thermodynamics of the dissociation of boric acid in synthetic seawater from 273.15 to 298.15 K. *Deep Sea Research Part A, Oceanographic Research Papers*, 37(5): 755-766. Doi: 10.1016/0198-0149(90)90004-F
- Dickson, A.G.; Sabine C.L. and Christian, J.R. (2007). Guide to best practices for ocean CO<sub>2</sub> measurements. Sidney, British Columbia, North Pacific Marine Science Organization, 176.
- Duarte, C.; Navarro, J.M.; Acuna, K.; Torres, R.; Manriquez, P.H. and Lardies, M. A. Vargas, C. A. *et al.* (2014). Combined effects of temperature and ocean acidification on the juvenile individuals of the mussel *Mytilus chilensis*. *Journal of Sea Research*, 85: 308–314. DOI: 10.1016/j.seares.2013.06.002.
- Dupont, S.; Dorey, N. and Thorndyke, M. (2010). What meta-analysis can tell us about vulnerability of marine biodiversity to ocean acidification? *Estuarine Coastal Shelf Science*. 89: 182-185. doi: 10.1016/j.ecss.2010.06.013
- El-Wazzan, E.; Abbas, A.S.; Abdel Razek, F.A. and Ragai, A. (2012). Reproductive Biology of the Carpet Shell Clam, *Tapes decussatus*, from Egyptian Coastal Waters. *Journal of Shellfish Research*, 31(1): 279 – 280. <https://doi.org/10.2983/035.031.0124>.
- FAO. (2018). The state of world Fisheries and Aquaculture production, Meeting the sustainable development goals, Food and Agriculture Organization of the United Nations, Rome. <http://www.fao.org/3/i9540en/I9540EN.pdf>
- Feely, R.A.; Sabine, C.L.; Lee, K.; Berelson, W.; Kleypas, J.; Fabry, V.J. and Millero, F.J. (2004). Impact of anthropogenic CO<sub>2</sub> on the CaCO<sub>3</sub> system in the oceans. *Science*, 305(5682): 362-366. doi:10.1126/science.1097329.
- Feely, R.A.; Sabine, C.L.; Byrne, R.H.; Millero, F.J.; Dickson, A.G.; Wanninkhof, R.; Murata, A.; Miler, L.A. and Greeley, D. (2012). Decadal changes in the aragonite and calcite saturation state of the Pacific Ocean, *Global Biogeochemical Cycles*, 26(3):1-15. doi:10.1029/2011GB004157.

- Fernandez-Reiriz, M.J.; Range, P.; Alvarez-Salgado, X.A. and Labarta, U. (2011). Physiological energetics of juvenile clams *Ruditapes decussatus* in a high CO<sub>2</sub> coastal ocean. *Marine Ecology Progress Series*, 433: 97–105. doi:10.3354/meps09062
- Filgueira, R.; Comeau, L.A.; Landry, T.; Grant, J.; Guyonnet, T. and Mallet, A. (2013). Bivalve condition index as an indicator of aquaculture intensity: A meta-analysis. *Ecological Indicators*, 25: 215–229. <https://doi.org/10.1016/j.ecolind.2012.10.001>
- Fernández-Reiriz, M.; Range, P.; Álvarez-Salgado, X.; Espinosa, J. and Labarta U. (2012). Tolerance of juvenile *Mytilus galloprovincialis* to experimental seawater acidification. *Marine Ecology Progress Series* 454: 65-74.
- Gazeau, F.; Quiblier, C.; Jansen, J.M.; Gattuso, J.P.; Middelburg, J.J. and Heip, C.H.R. (2007). Impact of elevated CO<sub>2</sub> on shellfish calcification. *Geophysical Research Letters*, 34(7): 1-5. doi:10.1029/2006GL028554
- Gazeau, F.; Alliouane, S.; Bock, CH.; Bramanti, L.; López Correa, M. and Gentile, M.; Hirse, T. *et al.* (2014). Impact of ocean acidification and warming on the Mediterranean mussel (*Mytilus galloprovincialis*). *Frontiers in Marine Science*, 62(1): 1-12. doi.org/10.3389/ fmars.2014.00062
- Grasshoff, K.; Erhardt, M. and Kremling, K. (1983). *Methods of Seawater, Analysis*. Verlag Chemi, Weinheim, New York, 419.
- Gruber, N.; Hauri, C.; Lachkar, Z.; Loher, D.; Frölicher, T.L. and Plattner, G.K. (2012). Rapid progression of ocean acidification in the California Current System. *Science*, 337(6091): 220-223. doi:10.1126/science.1216773.
- Hönisch, B.; Hemming, N.G.; Archer, D.; Siddall, M. and McManus, J.F. (2009). Atmospheric carbon dioxide concentration across the mid-Pleistocene transition. *Science*, 324(5934): 1551–1554. doi:10.1126/science.1171477.
- Ibrahim, H.A.H.; El-Sayed, W.M.M.; Shaltout, N.A. and El-Shorbagi, E. (2014). Effects of different pCO<sub>2</sub> concentrations on marine bacterial community structure, Eastern Harbor, Alexandria, Egypt, *Life Science Journal*, 11(10), 781-789. <http://www.lifesciencesite.com.126>
- Khairy, H.M.; Shaltout, N.A.; El-Naggar, M.F. and El-Naggar, N.A. (2014). Impact of Elevated CO<sub>2</sub> Concentrations on the Growth and Ultrastructure of Non-calcifying Marine Diatom (*Chaetoceros gracilis* F.Schütt), *The Egyptian Journal of Aquatic Research*, 40 (3): 243-250.
- Lewis, E. and Wallace, D.W.R. (1998). Program developed for CO<sub>2</sub> system calculations, Carbon Dioxide Information Analysis Center, Report ORNL/CDIAC-105, Oak Ridge National Laboratory Environmental Sciences Division, v. 4735, U.S. Department of Energy, Oak Ridge, Tennessee.
- Lucrezia, C.; Tommaso, S.; Antonietta, S.; Marisa, F. and Paolo, B. (2011). Bio indicators for siting the carpet clam *Tapes decussatus* L. farming in Mediterranean lagoons. *International Journal of Fisheries and Aquaculture*. 3(3): 53-63. Doi: <http://www.academicjournals.org/IJFA>
- Mehrbach, C.; Culberson, C.H.; Hawley, J.E. and Pytkowicz, R.N. (1973). Measurement of the apparent dissociation constants of carbonic acid in seawater at atmospheric pressure. *Limnology and Oceanography* 18: 897–907. <https://doi.org/10.4319/lo.1973.18.6.0897>.
- Melzner, F.; Stange, P.; Trübenbach, K.; Thomsen, J.; Casties, I.; Panknin, U. and Gorb, S.N. *et al.* (2011). Food Supply and seawater pCO<sub>2</sub> impact calcification and internal shell dissolution in the blue mussel *Mytilus edulis*. *PLoS One* 6 (9): 1-9. <https://doi.org/10.1371/journal.pone.0024223>

- Melatunan, S.; Calosi, P.; Rundle, S.; Widdicombe, S. and Moody, A. (2013). Effects of ocean acidification and elevated temperature on shell plasticity and its energetic basis in an intertidal gastropod. *Mar. Ecol. Prog. Ser.* 472: 155-168.
- Michaelidis, B.; Ouzounis, C.; Palaras, A. and Pörtner, H.O. (2005). Effects of long-term moderate hypercapnia on acid – base balance and growth rate in marine mussels *Mytilus galloprovincialis*. *Marine Ecology Progress Series*, 293: 109–118. Doi: 10.3354/meps293109.
- Navarro, J.M.; Torres, R.; Acuna, K.; Duarte, C.; Manriquez, P.H.; Lardies, M.; Lagos, N.A., *et al.* (2013). Impact of medium-term exposure to elevated pCO<sub>2</sub> levels on the physiological energetic of the mussel *Mytilus chilensis*. *Chemosphere*, 90(3): 1242-1248. <https://doi.org/10.1016/j.chemosphere.2012.09.063>.
- Orr, J.C. (2005). Anthropogenic ocean acidification over the twenty-first century and its impact on calcifying organisms. *Nature* 437: 681-686.
- Pan, T.C.F.; Applebaum, S.L. and Manahan, D.T. (2015). Experimental ocean acidification alters the allocation of metabolic energy. *Proceedings of the National Academy of Sci.*, 112(15): 4696-4701. <https://doi.org/10.1073/pnas.1416967112>.
- Pörtner, H.O., and Farrel, A.P. (2008). Physiology and climate change. *Science*, 322(5902): 690-692. DOI: 10.1126/science.1163156.
- Prado-Alvarez, M.; Gestal, C.; Novoa, B. and Figueras, A. (2009). Differentially expressed genes of the carpet shell clam *Ruditapes decussatus* against *Perkinsus olseni*. *Fish and Shellfish Immunology* 26: 72-83. [doi.org/10.1016/j.fsi.2008.03.002](https://doi.org/10.1016/j.fsi.2008.03.002).
- Range, P.; Chícharo, M.A.; Ben-Hamadou, R.; Piló, D.; Matias, D.; Joaquim, S.; Oliveira, A.P. *et al.* (2011). Calcification, growth and mortality of juvenile clams *Ruditapes decussatus* under increased pCO<sub>2</sub> and reduced pH: variable responses to ocean acidification at local scales? *Journal of Experimental Marine Biology and Ecology*. 396(2): 177-184. <https://doi.org/10.1016/j.jembe.2010.10.020>.
- Ringwood, A.H. and C.J. Keppler. (2002). Water quality variation and clam growth: is pH really a non-issue in estuaries? *Estuaries*. 25(5): 901–907. <https://doi.org/10.1007/BF02691338>.
- Sarazin, G.; Michard, G. and Prevot, F. (1999). A rapid and accurate spectroscopic method for alkalinity measurements in sea water samples. *Water Research*. 33(1): 290-294. [https://doi.org/10.1016/S0043-1354\(98\)00168-7](https://doi.org/10.1016/S0043-1354(98)00168-7).
- Stumpp, M.; Wren, J.; Melzner, F.; Thorndyke, M.C. and Dupont, S.T. (2011). CO<sub>2</sub> induced seawater acidification impacts sea urchin larval development I: Elevated metabolic rates decrease scope for growth and induce developmental delay. *Comparative Biochemistry and Physiology Part A: Molecular and Integrative Physiology*. 160(3): 331-340. <https://doi.org/10.1016/j.cbpa.2011.06.022>.
- Shirayama, Y. and Thornton, H. (2005). Effect of increased atmospheric CO<sub>2</sub> on shallow water marine benthos. *Journal of Geophysical Research*, 10(C09S08): 1-7. doi:10.1029/2004JC002618.
- Talmage, W.C. and Gobler, C.J. (2011). Effects of elevated temperature and carbon dioxide on the growth and survival of larvae and juveniles of three species of Northwest Atlantic bivalves. *PLoS ONE*, 6(10): 1-12 <https://doi.org/10.1371/journal.pone.0026941>.
- Thomsen, J. and Melzner, F. (2010). Moderate seawater acidification does not elicit long-term metabolic depression in the blue mussel *Mytilus edulis*. *Marine Biology*. Volume 157(12): 2667–2676. <https://doi.org/10.1007/s00227-010-1527-0>.

- Thomsen, J.; Casties, I.; Pansch, C.; Körtzinger, A. and Melzner, F. (2013). Food availability outweighs ocean acidification effects in juvenile *Mytilus edulis*: laboratory and field experiments. *Global Change Biology*, 19(4):1017–1027. doi:10.1111/gcb.12109
- Van Heuven, S.M.A.C.; Hoppema, M.; Huhn, O.; Slagter, H.A. and de Baar, H.J.W. (2011b). Direct observation of increasing CO<sub>2</sub> in the Weddell Gyre along the Prime Meridian during 19732008, Deep Sea Research Part II: Topical Studies in Oceanography, 58 (2526): 2613–2635, doi: 10.1016/j.dsr2.2011.08.007.
- Velasco, L.A. and Navarro, J.M. (2005). Feeding physiology of two bivalves under laboratory and field conditions in response to variable food concentrations. *Marine Ecology Progress Series*, 291: 115–124. doi:10.3354/meps291115.
- Walne, P.R. (1976). Experiments on the culture in the sea of the butterflyfish *Venerupis decussata* L. *Aquaculture* 8: 371–381. doi.org/10.1016/0044-8486(76)90119-8.
- Wang, Y.; Li, L.; Hu, M. and Lu, W. (2015). Physiological energetic of the thick shell mussel *Mytilus coruscus* exposed to seawater acidification and thermal stress. *Science of the Total Environment*, 1: 261–272. doi: 10.1016/j.scitotenv.2015.01.092.
- Wood, H.L.; Spicer, J.I. and Widdicombe, S. (2008). Ocean acidification may increase calcification rates, but at a cost. *Proceedings of the Royal Society B: Biological Sciences*, 275(1644): 1767–1773. <http://doi.org/10.1098/rspb.2008.0343>.