



## Evaluation of Water Quality and Plankton for Mahmoudia Canal in Northern West of Egypt

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### ABSTRACT

Water quality describes the condition of the water, including chemical, physical, and biological characteristics, usually with regarding to its suitability for a particular purpose. The present study describes the environmental conditions of Mahmoudia canal which are the main water source for Alexandria and Behiera governorates. Water samples were collected from ten sites along the Mahmoudia canal in the year 2018 for study the seasonal variations of the water quality parameters (physicochemical and biological parameters). The measured physicochemical parameters include; (Turbidity, Temperature, pH, Electrical Conductivity, Total Dissolved Solids, Total Suspended Solids, Total alkalinity, Total Hardness, Nitrate, Nitrite, Ammonia, Sulfate, Phosphate, Dissolved Oxygen, Biological Oxygen Demand, Chemical Oxygen Demand, Chlorinated pesticides and some heavy metals). Phytoplankton and Zooplankton were examined as biological parameters. The values of average water quality index (WQI) of Mahmoudia canal ranged from 75.01 (poor water quality) to 132.57 (very poor water quality). According to the results of Cluster analysis, six statistically significant clusters were formed. Three groups of phytoplankton were recorded which, the annual maximum density of phytoplankton recorded in Town of Abou Hommos, while the lowest recorded in site Zawyet Ghazal Village. The highest density registered in winter (3460.2 U/ml), while the lowest density recorded in summer (831.6 U/ml). In addition to six zooplankton groups were quantified in the analysis of the samples from ten sites, the highest density listed in summer (6000 ind./m<sup>3</sup>), although the lowest listed in winter (3200 ind./m<sup>3</sup>).

### INTRODUCTION

Water is super abundant on the planet as a whole, but fresh potable water is not always available at the right time or the right place for human or ecosystem use. Water quality refers to the physical, chemical and biological characteristics of water. It is a measure of the condition of water relative to the requirements of one or more biotic species and or to any human need or purpose. Rivers are vital and vulnerable freshwater ecosystems that are critical for the sustenance of all life. Rivers are waterways of strategic importance across the world, providing main water resources

for domestic, industrial and agricultural purposes (Aggarwal and Arora, 2012). Degradation of water quality is a major issue in Egypt.

In spite of the Egyptian Environmental Affairs Agency (EEAA) and the ministry of the environmental protection legislation's, some institutions and drains still discharge their pollutants into the river Nile. These pollutants adversely change the environment in the river by changing the growth rate of species and interfere with the food chain or with public health (Shaaban *et al.*, 2004). Spearman Rank correlation analyses were performed on environmental parameters and zooplankton groups to examine significant relationships. Application of the WQI is also suggested as a very helpful tool that enables the public and decision makers to estimate water quality of fish farms (Heneash, 2015). Salinity, dissolved oxygen and pH may be responsible for the variations in zooplankton community structure (Heneash *et al.*, 2015).

The severity of present water quality problems in Egypt varies among different water bodies depending on: flow, use pattern, population density, the extent of industrialization, availability of sanitation systems and the social and economic conditions existing in the area of the water source. Discharge of untreated or partially treated industrial and domestic wastewater, leaching of pesticides and residues of fertilizers; and navigation are often factors that affect the quality of water (EEAA, 2002). Mahmoudia canal can be readily contaminated by human activity without any obvious signs. The constituents of concern in domestic and municipal wastewater are pathogens, parasites, nutrients, Oxygen demanding compounds and suspended solids. In Greater Cairo and other cities, the sewerage systems also serve industrial and commercial activities. As these toxic substances (heavy metals, persistent, organic micro pollutants and abundant) are mainly attached to the suspended material, most of it collects in the sludge that can accumulate in aquatic habitats and their concentration increases through biomagnification. Improper sludge disposal or reuse may lead to contamination of surface and ground water (EEAA, 2002). Few studies were published about Mahmoudia canal, (Abdullah and Hussona, 2014) showed that the main water quality is exhibit high pollution levels that create health risks at present, indicates unsafe levels of pollution for direct use in drinking water, irrigation and fisheries.

Therefore, the present study was carried out to monitor the changes in water quality at ten sites in Mahmoudia canal, Egypt, which are a useful source of information for the understanding of seasonal and relationship between physical, chemical and biological changes in Aquatic environments in Mahmoudia canal water quality that it is the main water sources for Alexandria city.

## MATERIALS AND METHODS

### Study area

Mahmoudia canal locates at northern west of Egypt at the northern edge of the Beheira Governorate (receive about 15 Mm<sup>3</sup>/day). It receives water from the Rosetta branch in Mahmoudia city at km 194.200 with actual served area of 130,200 hectares. (Abo Kila, 2012).

The total length of the canal is 77.170 km and there are seventy canals off-take from this canal. Al-Mahmoudia canal has three sources of water; two fresh water sources which are from Rosetta branch via El-Atf pump stations at the head of the canal, and Al-Khandaq Eastern canal at km 13.200 on Al-Mahmoudia canal, the third is drainage water from Zarkon drain at km 8.500 on Al-Mahmoudia canal via Edko irrigation pump station which lifting part of Zarkon drain water into Al-Mahmoudia

canal. The canal receives domestic and agriculture wastes from Zarkon drain and other non-point sources.

### Sampling, measuring and analysis

Water samples were collected seasonally through winter, spring, summer and autumn of 2018. Ten stations were chosen represented different environments of the canal for the present study. The locations description of the sampling stations is shown in Figure (1).

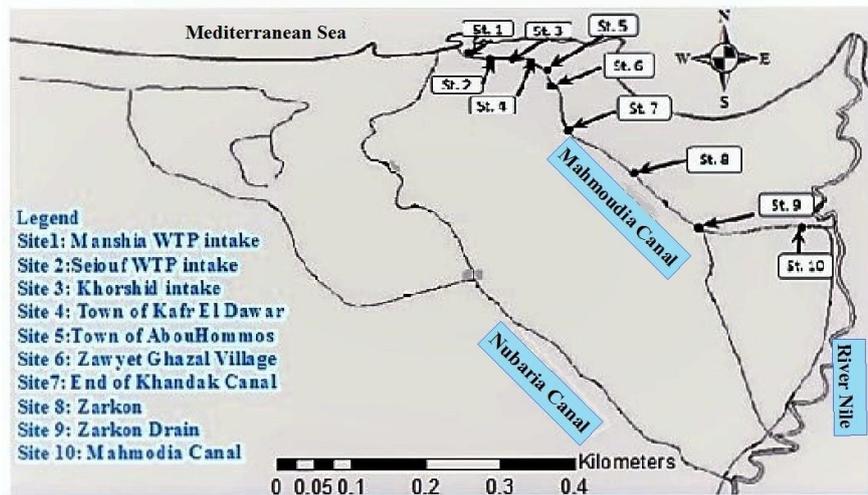


Fig. 1: Location map of Al-Mahmodia canal and sampling sites.

### Physicochemical analysis

Measuring and analysis were done upon physical, chemical, microbiological parameters. These parameters were measured seasonally (Table 1). Temperature and pH parameters were estimated at the spot immediately after the collection of the samples. Analyses of water samples were performed using standard analytical methods according to procedures outlined in the standard methods for (APHA, 2005).

Table 1: Measuring methods of physical and chemical parameters.

Parameter	Symbol	Units	Test Method
pH			Digital electrode pH meter
Turbidity	Turb.	NTU	Nephelometer
Temperature	T		an analytical thermometer
Electrical Conductivity at 25°C	EC	μS/cm	Digital conductivity meter
Biochemical Oxygen Demand	BOD	mg/l	Azide modification of winkler's titration method and determined by using 5 days method
Chemical Oxygen Demand	COD	mg/l	Reflux titrimetry method
Total Dissolved Solids at 105°C	TDS	mg/l	filtration a volume of sample with glass micro fiber filter (GF/C) and a known volume of filtrate was evaporated
Total Suspended Solids	TSS	mg/l	Membrane filtration method
Total Hardness	TH	mg/l	EDTA complexometric titration method using Erichrome black-T dye as indicator
Total Alkalinity	T Alk.	mg/l	by titration with standard sulphuric acid using phenolphthalein and methyl orange as indicator (pH = 4.5)
Nitrate	NO <sub>3</sub>	mg/l	Acid treatment followed by spectrophotometry
Nitrite	NO <sub>2</sub>	mg/l	
Ammonia	HH <sub>3</sub>	mg/l	Distillation titrimetric method
Chloride	Cl	mg/l	Argentometric titration
Sulphate	SO <sub>4</sub>	mg/l	Turbidimetric methods
Phosphate	PO <sub>4</sub>	mg/l	Ascorbic acid spectrophotometry
Pesticides			Chromatographic techniques ( EPA 508.1, 537.1 and 508.1 methods)

### Bacteriological and Biological analysis

Samples for bacteriological analyses were collected into sterilized plain glass bottles. The most accurate measurements of water quality are made on-site because water exists in equilibrium with its surroundings. The samples were kept in ice on the field and after that refrigerated in the laboratory. Total coliform (TC, CFU/100ml) and fecal coliform (FC, CFU/100ml) were carried out according to standard methods for water and wastewater (APHA, 2005). Water samples were collected using a Rüttner bottle, and estimation of the phytoplankton standing crop was carried out by sedimentation method as reported in a standard method in American Public Health Association (APHA, 1999), the samples were preserved in a solution of 4% formaldehyde. Phytoplankton samples were examined in the laboratory using a Research microscope, identification to species and counting will do, the species counts were expressed as unit's  $l^{-1}$ .

The zooplankton samples were collected by filtering 50 liters of water through a small standard plankton net (mesh size 55  $\mu m$ ) using a plastic container of liters' capacity. The collected samples were preserved directly with 4% neutral formalin solution in 250 ml polyethylene bottles. The volume of each sample was concentrated to 100 ml and the whole sample was examined in a Petri dish under a research binocular microscope. For zooplankton enumeration purposes, at least two aliquots (2 ml of well-shaken suspension) were withdrawn from each sample using a graduated pipette, placed in a counting chamber and the number of individuals of each species was counted. The average number of duplicated amination for each sample was estimated and the counts were expressed as number of individuals per cubic meter (Heneash, 2015).

### Water Quality Index (WQI)

WQI is a mathematical approach to summarize multiple properties into a single value. WQI is defined as a rating reflecting the composite influence of different water quality parameters on the overall quality of water. The WQI was computed through three steps. First, each of the 12 parameters (Turbidity, pH, TDS, COD, BOD, total hardness, total alkalinity Cl,  $NO_3$ ,  $NO_2$ ,  $SO_4$ ) was assigned a weight ( $w_i$ ) according to its relative importance in the overall quality of water for drinking purposes (Table 2). The maximum weight 5 was assigned to pH, COD and nitrate because of its major importance in water quality assessment; other parameters were assigned weights between 1 and 5 based on their relative significance in the water quality evaluation. Second, the relative weight ( $W_i$ ) of the chemical parameter was computed using the following equation.  $W_i = w_i / \sum_{i=1}^n w_i$

Where:  $W_i$  is the relative weight,  $w_i$  is the weight of each parameter, and  $n$  is the number of parameters. Calculated relative weight ( $W_i$ ) values of each parameter are given in Table (2). In the third step, a quality rating scale ( $q_i$ ) for each parameter is assigned by dividing it's the concentration in each water sample by its respective standard according to guidelines (EWQS, 2007), and the result is multiplied by 100:

$$q_i = (C_i/S_i) * 100$$

Where:  $q_i$  is the quality rating,  $C_i$  is the concentration of each chemical parameter in each water sample in mg/L, and  $S_i$  is the Egyptian drinking water standard for each chemical parameter in mg/L. For computing WQI, the sub index (SI) is first determined for each chemical parameter, as given below:  $SI_i = W_i \times q_i$   
 $WQI = \sum SI_i - n$ ; Where:  $SI_i$  is the sub index of  $I^{th}$  parameter;  $W_i$  is relative weight of  $I^{th}$  parameter;  $q_i$  is the rating based on concentration of  $I^{th}$  parameter, and  $n$  is the number of chemical parameters. The standard allows able value ( $S_i$ ) for each parameter was taken from two criteria Egyptian drinking water quality standards, (EWQS, 2007). The

computed WQI values are classified into five categories: Excellent water (WQI < 50); good water (WQI=50–100); poor water (WQI=100–200); very poor water (WQI=200–300); and water unsuitable for drinking (WQI > 300) (Batabyal and Chakraborty, 2015).

**Table 2: Relative weight of chemical parameters.**

Chemical parameters <sup>a</sup>	Egyptian Standard <sup>b</sup>	Weight (wi)	Relative weight (Wi)
<b>PH</b>	<b>7.5</b>	5	0.108
<b>Turbidity</b>	<b>1</b>	4	0.086
<b>Total Dissolved solids at 105°C</b>	<b>1000</b>	4	0.086
<b>Ammonia, free</b>	<b>0.5</b>	4	0.086
<b>Chemical Oxygen Demand (COD)</b>	<b>6</b>	5	0.108
<b>Biological Oxygen Demand (BOD)</b>	<b>10</b>	4	0.086
<b>Hardness, Total</b>	<b>500</b>	3	0.065
<b>Alkalinity, Total</b>	<b>100</b>	4	0.086
<b>Chloride</b>	<b>250</b>	3	0.065
<b>Nitrate</b>	<b>45</b>	2	0.043
<b>Nitrite</b>	<b>0.2</b>	5	0.108
<b>Sulphate</b>	<b>250</b>	3	0.065
		$\sum w_i = 46$	$\sum W_i = 1$

(a) Chemical parameters in mg/L. (b) Values indicates desirable limit, in absence of alternate source (EWQS, 2007).

### Statistical analysis

Statistical analysis was computed by Microsoft office Excel 2010. Inter relationships between different parameters were determined from the correlation matrix formed in IBM SPSS statistics, ver. 22 program. The correlation coefficients are considered significant at the 95% confidence level ( $p \leq 0.05$ ) and 0.01 levels (2-tailed). Cluster Analysis (CA) was applied to the results by using the MVSP (version 3.1) package program. Cluster analysis was done using Ward's method (z-transformation of the input data, Euclidean distance as similarity measure) based on the standardized mean of 16 measured physicochemical water parameters obtained from the four seasons as Turbidity, pH, EC, TDS, TSS, NH<sub>3</sub>, COD, BOD, hardness, T.Alkalinity, Cl, PO<sub>4</sub>, NO<sub>3</sub>, NO<sub>2</sub>, Si and SO<sub>4</sub> were used as variables to show the spatial heterogeneity among the stations.

## RESULTS AND DISCUSSION

### Physio-chemical and biological parameters

Seasonal variations of physicochemical and biological parameters of the water samples from the ten sampling sites on the Mahmoudia canal were performed as shown in Table (3).

The present study revealed that, pH values were slightly higher during the winter compared to the other seasons and it was accompanied by increase in dissolved oxygen. The pH value of water is controlled by the dissolved oxygen, water temperature, sewage discharge, decomposition of organic matter and photo synthetic activities (Nassar and Hamed, 2003). pH is not only a measure of potential pollutants but also is related to the concentration of many other substances, particularly the weakly dissociated acids and bases (Abdelmongy, and El-Moselhy, 2015).

Electrical conductivity values varied between 443 and 2056  $\mu\text{S}/\text{cm}$ . It's maximum value recorded in the winter at all sampling sites 11084  $\mu\text{S}/\text{cm}$  and the

minimum value was in the summer in all sampling fields 5085  $\mu\text{S}/\text{cm}$  which may be attributed to silts which it brings during its course in the rainy season. The average value of typical, unpolluted rivers is approximately 350  $\mu\text{S}/\text{cm}$  (Koning and Roos 1999). Therefore, the parameter does not give cause for concern and it makes the water suitable for direct domestic use.

Table 3: Average of physical, chemical and biological analysis of different stations in the study area.

S. No	Turb.	PH	EC	TDS	TSS	NH <sub>4</sub>	COD	BOD	Hardness	T. Alk	Cl	PO <sub>4</sub>	NO <sub>3</sub>	NO <sub>2</sub>	Si	SO <sub>4</sub>	Temp	Fecal.coliform	T.coliform	Heterotrophic plate
1	9.3	7.7	624.2	628	12	4.6	23.9	7.7	176.2	195	60.3	0.2	5.3	0.3	2.95	61.5	20.7	250.2	5200	6200
2	4.4	7.7	788	487.7	8.6	3	25.7	10.7	218	221	58.7	0.06	10.6	0.6	1.7	95.9	20	153	5400	10000
3	14.3	7.8	638	625.5	12.6	3.0	19.1	17	173	199	58.3	0.2	5.2	0.3	3.9	70.2	20.3	601	5175	6550
4	8.4	7.8	504	413	10	2.9	21	3.8	163	171	62.3	0.17	4.3	0.2	1.0	44.1	20.3	352	10175	4400
5	9.6	7.8	609	553	10	2.4	20.7	16	166	187	79.3	0.1	5.6	0.4	1.8	56.6	20	1167	6225	16675
6	7.9	7.8	911	359	11.3	2.3	20	2.5	249	209	139	0.09	8.4	0.5	1.3	170.9	20	1602	9500	6625
7	6.1	7.7	776	356	7.33	2.4	24	9.7	181	196	85	0.1	7.3	0.5	2.9	128	19	1700	9500	19450
8	5.6	7	988	351	12	2.6	36	16	232	210	143	0.1	10	0.6	3.9	355	19	2475	26700	64350
9	6.8	7.7	762	451	10	2.2	42	9.7	182	196	81.9	0.12	7.6	0.4	3.2	113	20	1500	8275	15000
10	7.1	7.6	971	342	12	2	37	12	219	203	135	0.06	10	0.4	3.3	228	59	2633	5350	10425

Turbidity values were ranged from 2.71 to 20.8 NTU. The maximum values were in the winter with 20.8 NTU while the minimum value was 2.71 NTU recorded in the summer. The reason for the recorded higher turbidity in winter season may be due to the mixing of sewage water and high rate of evaporation.

The results of the present study revealed that the maximum value of biological oxygen demand (BOD) was observed in the winter with 69.2 mg/L and the minimum values in the summer was 0.9 mg/L in all sampling sites. The highest value of BOD was above 10 mg/L which showed in some station could be attributed to the abundance of bacterial and microbial activity in these areas and/or the decomposition of organic matter, this agrees with (Abdar, 2013).

Chemical oxygen demand (COD) is a reliable parameter for judging the extent of pollution in water. The highest value was noted 54 mg/L in autumn season and minimum value was 6.4 mg/L in the summer season. A trend of increasing COD level showing the population load and activities caused by the mixing of sewage water, garbage dumping and industrial discharges. This increase in COD indicates that the pollution may increase and hence need exercise of monitoring, abatement and control.

Nutrient salts: Nitrates levels varying from 3.23 mg/L to 17.99 mg/L which the lowest nitrate concentration was 5.2 - 8.26 during spring, while the highest nitrate concentrations were recorded during winter at the different sites may be owing to in the winter mixing deepening increased nitrate enrichment into the euphotic zone from deeper water (Al-Qutob *et al.*, 2002), and the lower concentrations during the other seasons may due to high flourishing of phytoplankton observed (Ferrer, 1996).

Concentrations of dissolved nitrite in water in the present study sites ranged from 0.05-1.49 mg/L. The highest nitrite concentrations recorded in the summer, while the lowest concentrations recorded during the autumn season.

Ammonia concentrations ranged from 0.14-8.5 mg/L, the highest ammonia concentration recorded in autumn and spring season in while the lowest concentration recorded during the winter season. Light has an inhibitory effect on ammonium oxidation by nitrifying bacteria (Guerrero and Jones, 1996). Presence of ammonia is an indicator of pollutants of high activity and related to the presence of wastewater treatment plant and industrial discharges near this location. The excess usage of inorganic fertilizers containing nitrogen is the main source of nitrate ions in the water

sources. Also, the complete oxidation of nitrogen from the decaying of organic materials in the surface water and the sediment is an important factor in producing ammonia, nitrate and nitrite (Taha *et al.*, 2004).

The concentration of available  $\text{PO}_4^{3-}$  varied from 0.02 to 0.56 mg/L, respectively with higher levels in the summer. Low phosphate concentrations in the examination, may be due to uptake by phytoplankton and primary producers (Al-Qutob *et al.*, 2002).

**Anions:** The Chloride contents were ranged from 28.6 to 374 mg/L, which chloride content was lower than the accepted limit of 250 mg/L at the sampling sites in all season of the canal except 6, 8 and 10 stations. The levels winter were higher than those obtained in other seasons and the minimum levels were in the summer season. The lowest concentration of chloride was affected more with fresh water from different drains.

Sulfate concentration in the canal varied from 29 to 1227 mg/L. The concentration of  $\text{SO}_4^{2-}$  was lower in the summer as compared to other seasons and the maximum levels recorded in the winter season which might be contributed by agricultural runoffs. From the observations, it can be seen that sulfate was present under acceptable limits expect in the winter season.

The anions and cations are naturally very variable in surface waters due to climatic and geographical conditions also; their distribution depends on the evaporation rate and the drainage water from different drains (Shama *et al.*, 2011).

**Hardness:** Total hardness gives information about the concentration of Ca and Mg ions. The hardness of the canal water fluctuated between 138.8 and 456.8 mg/L. The trend of variation was non-uniform in all the seasons but it increased in winter season and decreased in summer season.

In the present investigation, total Alkalinity ranged between 158 and 276 mg/L. Alkalinity at all the seasons in all the sites was above the desirable limit prescribed for drinking water which is 120 mg/L (WHO). The high values of Alkalinity may also be due to an increase in free carbon dioxide in the river which ultimately results in the increase in Alkalinity. High Alkalinity may cause problems if water is used for irrigation purposes as high alkalinity leads to increase in relative proportion of sodium in soil by precipitating Ca and Mg ions.

**Total dissolved solids** and **total suspended** are common indicators of polluted waters. TDS values ranged from 266 to 1234 mg/L, and the results were high in winter and low in summer. TSS values ranged from 6 to 18 mg/L. The values were low as compared with the (IS) Standards therefore the palatability of water with TDS and TSS can be considered to be good. At all the sites in completely seasons it was observed that the dissolved solid contents were found to be greater than the suspended solid levels may be due to is indicative of erosion that has taken place during the river.

**Chlorinated Pesticides:** The concentration of pesticides such as a- BHC, b- BHC, g- BHC (Lindane), d-BHC, Heptachlor, Heptachlor epoxide(isomer), g-Chlordane and Endosulfan I, were undetectable indicating that the canal is free from them and hence safe for drinking purposes. Galal, (1983) reported that Lindane, DDE and DDT were found in Mahmoudia canal and the concentrations were ranged from 0.91 to 2.32 µg/l.

**Heavy metals:** in this study some heavy metals were investigated in the autumn season as shown in Figure (2). Silver (Ag), Boron (B), Beryllium (Be), Copper (Cu) and Cadmium (Cd) were undetectable at both the sites, in addition to Nickel (Ni), and Lead (Pb) at most sites therefore the water of canal is free from toxic metals.

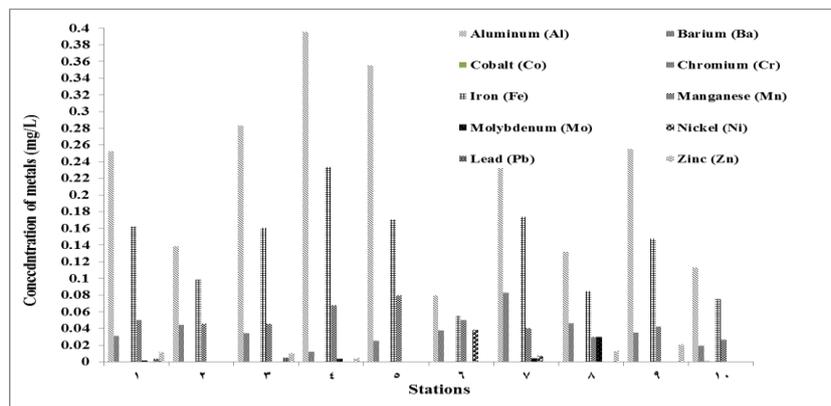


Fig. 2: Concentration of heavy metals in ppm from Mahmoudia canal.

Hence there is no danger with respect to these metals Barium (Ba), Cobalt (Co), Chromium (Cr), Iron (Fe), Manganese (Mn), Molybdenum (Mo), Nickel (Ni), Lead (Pb) and Zinc (Zn) were also below the detection limits.

The concentrations of Aluminum (Al) in the water varied from 0.08 to 0.395 mg/L. Both towns of Kafr El Dawar and Abou Hommos sites exceeded the background level with values 0.395 and 0.355 mg/L, respectively and the WHO limit of 0.2 mg/L. Aluminum being non-toxic presents no health hazards at concentrations normally found in natural waters, but at very high aluminum levels can sometimes cause water to have a bluish color. Galal, (1983) also cited that the heavy metals (Cu, Pb, Zn, Mn and Fe) were found in the canal at levels less than the maximum permissible limits.

## Biological parameters

### Phytoplankton:

Table 4 shows that, the average three groups of phytoplankton. The lowest listed in summer (831.6 U/ml) because zooplankton feeds on phytoplankton.

Table 4: Seasonal variation of phytoplankton (U/ml) at Al-Mahmodia canal in 2018

	Winter										Average
	1	2	3	4	5	6	7	8	9	10	
Phytoplankton	4185	502	3761	3297	13841	832	1901	1731	2970	1582	3460.2
Blue Green Algae	92	61	92	246	366	58	161	358	266	415	211.5
Green Algae	341	51	216	225	929	116	311	440	195	284	310.8
Diatoms	3553	276	3395	2642	12292	541	1187	922	2058	860	2772.6
	Spring										
Phytoplankton	1013	1322	743	2033	1962	857	1107	1335	2611	1660	1464.3
Blue Green Algae	133	200	44	22	411	121	198	108	294	97	162.8
Green Algae	214	420	133	282	555	267	262	206	403	333	307.5
Diatoms	457	318	330	990	711	515	544	726	992	917	650
	Summer										
Phytoplankton	983	991	492	1464	502	664	547	806	1487	380	831.6
Blue Green Algae	184	265	82	338	90	163	58	64	143	58	144.5
Green Algae	553	380	133	605	260	359	336	512	762	184	408.4
Diatoms	184	173	236	352	125	131	134	205	493	81	211.4
	Autumn										
Phytoplankton	1042	1382	740	2035	1962	859	1107	1335	2611	1669	1474.2
Blue Green Algae	134	205	33	0	493	70	198	108	294	92	162.7
Green Algae	218	420	133	288	556	268	262	206	403	328	308.2
Diatoms	465	317	333	998	744	514	544	726	992	911	654.4
	Average										
Blue Green Algae	108.6	146.2	50.2	121.2	272	82.4	123	127.6	199.4	132.4	136.3
Green Algae	331.5	317.75	153.75	350	575	252.5	292.75	341	440.75	282.25	333.725
Diatoms	1164.75	271	1073.5	1245.5	3468	425.25	602.25	644.75	1133.75	692.25	1072.1

The annual highest density recorded in site 5, while the lowest recorded in site 6. The highest density listed in winter (3460.2 U/ml), while the lowest density recorded in summer (831.6 U/ml) due to zooplankton feed on phytoplankton. Other hand the diatoms was the largest density of phytoplankton this result agrees with Heneash *et al.* (2015), while the lowest was Blue Green Algae.

### Zooplankton:

From the analyzed data in Table 5, a visible change in zooplankton community with regard to numerical abundance and composition was evident among sites and in the seasonal cycle. In average, 6 zooplankton groups were quantified through the analysis of the samples collected from ten sites in 4 seasons, the highest density listed in summer (6000 ind./m<sup>3</sup>) this results agree with Aboul Ezz *et al.* (2014), while the lowest listed in winter (3200 ind./m<sup>3</sup>) due to reverse relation between phytoplankton and zooplankton. The annual highest density recorded in sites 4 and 5, while the lowest recorded in site 9. Maybe, these organisms effects by light. The influence of variable light source intensity on the dial vertical migration of the zooplankton continues to be the subject of numerous studies (Martynova and Gordeeva, 2010). Also, a significantly reduced attraction to visible light was noted for the juvenile, Copepoda and Rotifera forms (Springer and Skrzypczak, 2015).

**Table 5: seasonal variation of zooplankton (ind./m<sup>3</sup>) at Al-Mahmodia canal in 2018**

Winter											
Groups	1	2	3	4	5	6	7	8	9	10	Average
Rotifers	1000	1000	0	1000	0	1000	0	1000	0	1000	600
Copepoda	1000	1000	1000	0	1000	1000	1000	0	0	1000	700
Protozoa	0	0	0	1000	1000	1000	1000	0	1000	0	500
Ostracoda	1000	0	0	1000	0	0	0	1000	0	2000	500
Nematoda	0	1000	1000	0	0	0	0	1000	1000	1000	500
Chordate	0	0	0	2000	1000	0	1000	0	0	0	400
Average	500	500	333.333	833.333	500	500	500	500	333.333	833.333	3200
Spring											
Groups	1	2	3	4	5	6	7	8	9	10	Average
Rotifers	0	2000	1000	1000	2000	1000	0	0	1000	1000	900
Copepoda	1000	0	0	1000	1000	1000	1000	2000	0	0	700
Protozoa	0	1000	1000	1000	1000	0	2000	0	0	1000	700
Ostracoda	1000	1000	0	2000	1000	0	0	2000	0	1000	800
Nematoda	2000	0	1000	1000	2000	0	0	1000	0	0	700
Chordate	1000	0	1000	0	2000	1000	0	1000	2000	1000	900
Average	833.333	666.667	666.667	1000	1500	500	500	1000	500	666.667	783.333
Summer											
Groups	1	2	3	4	5	6	7	8	9	10	Average
Rotifers	1000	2000	2000	1000	2000	2000	0	1000	1000	1000	1300
Copepoda	1000	1000	1000	1000	1000	1000	1000	0	0	0	700
Protozoa	0	0	0	2000	2000	2000	3000	2000	0	0	1100
Ostracoda	1000	2000	2000	2000	0	0	0	3000	0	2000	1200
Nematoda	2000	2000	2000	2000	0	0	0	1000	1000	1000	1100
Chordate	0	0	0	2000	2000	2000	0	0	0	0	600
Average	833.333	1166.67	1166.67	1666.67	1166.67	1166.67	666.667	1166.67	333.333	666.667	1000
Autumn											
Groups	1	2	3	4	5	6	7	8	9	10	Average
Rotifers	0	0	0	1000	2000	1000	0	0	1000	1000	600
Copepoda	0	0	0	1000	1000	1000	1000	0	0	0	400
Protozoa	0	1000	1000	1000	1000	0	0	0	0	1000	500
Ostracoda	1000	1000	0	2000	1000	0	0	2000	0	1000	800
Nematoda	2000	0	1000	1000	2000	0	0	1000	0	0	700
Chordate	1000	2000	0	1000	2000	1000	0	0	2000	1000	1000
Average	666.667	666.667	333.333	1166.67	1500	500	166.667	500	500	666.667	666.667

### Water Quality Index

The results of WQI and water type based on eleven parameters are presented in Table (6). The values of average water quality index (WQI) of Mahmoudia canal ranged from 75.01 (poor water quality) to 132.57 (very poor water quality). WQI registered the lowest value at site (4) during summer poor water quality, while other sites were very poor, in addition to maximum value was observed at site (1) during autumn.

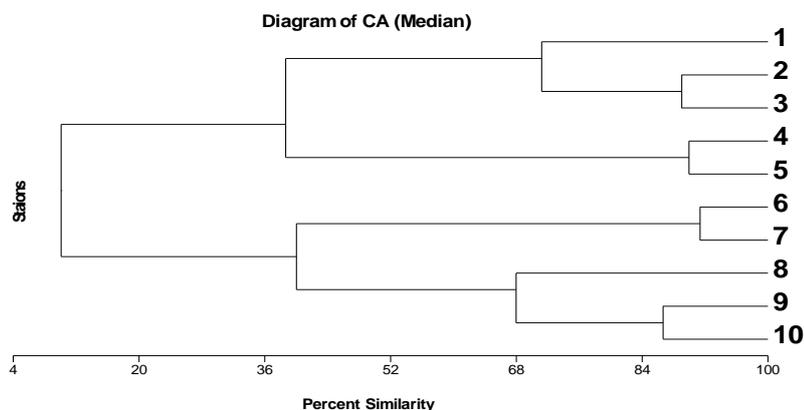
The quality of water in summer and winter season is poor, while in autumn and spring season move towards very poor conditions. This may be due to the water quality of which was influenced by organic and inorganic nutrient pollutants. The trend of the aquatic WQI was not clear due to the fluctuating nature of water quality. The discharge of return irrigation water, industrial and municipal wastes into canal water through drains, containing high levels of nutrient salts and heavy metals deteriorate the water quality of canal and accordingly in this period. In winter may dilution from rainy water reduce the activity of microorganisms, increase in oxygen and dilution to wastewater occurs.

**Table 6: WQI of Mahmoudia canal at different sampling points.**

Season	Site									
	1	2	3	4	5	6	7	8	9	10
Winter	40.489	111.07	126.86	99.59	36.77	40.08	74.54	175.03	102.83	131.35
Spring	210.35	102.5	160.4	96.64	156.21	139.5	122.93	124.33	145.35	129.18
Summer	55.46	18.45	43.52	8.47	50.70	56.56	46.9	90.64	57.72	82.75
Autumn	209.24	97.06	199.52	95.37	161.44	139.49	90.46	115.75	127.96	86.06
Average	<b>128.88</b>	<b>82.27</b>	<b>132.57</b>	<b>75.01</b>	<b>101.28</b>	<b>93.90</b>	<b>83.70</b>	<b>126.44</b>	<b>108.46</b>	<b>107.33</b>

### Cluster Analysis (spatial similarity and site grouping)

Cluster analysis is an unsupervised multivariate statistical technique used to classify the objects into clusters based on their similarity and it is one of the most widely used multivariate statistical technique to evaluate the surface water quality and it is typically illustrated by a dendrogram (Tokath *et al.*, 2013). Data were then amalgamated into dendrogram plots and the similarity coefficients of all the investigated stations on the Mahmoudia canals base on their water quality similarities are given in Figure (3).



**Fig. 3: Dendrogram viewing the relationship amongst sampling locations in Mahmoudia canal.**

According to the results of CA, six statistically significant clusters were formed. Cluster 1 (C1) corresponded to the station 1 was on the canal from Manshia WTP intake. Cluster 2 (C2) comprised stations of 2 and 3 that were on the canal from Seiouf WTP and Khorshid intake represented that site 2 and 3 have the same characteristics, while the cluster 3 (C3) corresponded to the stations of 4, and 5 that were for upstream of the Kafr El Dawar and Abou Hommos city, which have similar properties. The cluster 4 (C4) consisted from the station of 6 and 7 that related to the irrigation canal for Zawyet Ghazal village and end of Khandak canal and the changes in water quality in them were mainly. Cluster 5 (C5) which is related to station 8 that was close to the input of Zarkon. Cluster 6 (C6) corresponded to the stations of 9 and 10 that were on the irrigation canal of Mahmodia and Zarkon drain and the water quality parameter from this group was similar. It's obvious that group of cluster C1 to C3 in addition to C4 to C6 were similar characteristics, therefore the differences between the groups indicate the differences in the sources of pollution.

### Inter relationships

From the correlation matrix between various parameters, most of the parameters were found to statistically significant correlation with each other indicating close association of these parameters with each other as presented in Table (7). The conductivity of the water however, showed a highly positive correlation with hardness ( $r = 0.884$ ), sulfate ( $r=0.889$ ) and nitrate ( $r=0.897$ ) in addition to the discrepancy observed in the relationship between EC and TDS ( $r=-0.668$ ), nitrite ( $r=0.780$ ), total alkalinity ( $r=0.747$ ) and chloride ( $r = -0.863$ ). This might be attributed to their documented conductivity effect.

Total Alkalinity showed high significant positive relationship with conductivity ( $r =0.747$ ), nitrate ( $r =0.822$ ), total hardness ( $r =0.785$ ) and nitrite ( $r =0.852$ ). Total hardness showed high significant positive relationship with Chloride ( $r = 0.786$ ), conductivity ( $r =0.884$ ), total Alkalinity ( $r =0.785$ ), nitrite  $=0.709$ ), nitrate ( $r =0.828$ ) and sulfate ( $r = 0.741$ ). Similar conclusion was drowned out by Shinde *et al.*, (2011). Increasing concentrations of chlorides, mostly sodium chloride, increases Salinity of water. Not unexpectedly chloride concentrations hence, maintained strong correlations with conductivity ( $r =0.863$ ), hardness ( $r =0.786$ ), nitrate ( $r =0.628$ ), TDS ( $r = -0.738$ ) and sulfate ( $r = 0.864$ ).

Statistical analysis showed high positive correlations of nitrate with other several pollution attributes, such as: hardness ( $r = 0.828$ ), EC ( $r = 0.897$ ), sulfate ( $r= 0.728$ ), chloride( $r=0.862$ ), nitrite ( $r=0.842$ ), total alkalinity ( $r = 0.822$ ). These significant correlations indicated the impact of sewage discharge and agricultural runoff. Feacel caliform affected positively with nitrite ( $r=0.756$ ), nitrate ( $r=0.882$ ), EC ( $r = 0.919$ ), sulfate ( $r= 0.846$ ), in addition to that heterotrophic have high positive correlation with total coliform and sulfate ( $r=0.922$  and  $r=0.772$ , respectively) at  $p <0.001$ . Also, diatom recorded positive correlation with blue green and green algae ( $r=0.772$  and  $r=0.751$ , respectively at  $p < 0.05$ ) while green algae recorded high a significant positive correlation with pH ( $r=0.960$  at  $p < 0.01$ ).

The correlations of the Nematoda species were significant through the parameter;  $SO_4$ ,  $NO_3$ , EC, Cl, Si and Feacel Coliform with correlation ( $r =0.749$ ,  $r=0.641$ ,  $0.835$ ,  $r=0.681$ ,  $r=0.681$  and  $0.717$ ) respectively, while protozoa species showed negative correlation with TSS ( $r=-0.668$ ). There high negative significant correlation between nitrate and pH with ( $r= -0.726$ ), moreover between phosphate and conductivity ( $r = -0.662$ ), Hardness ( $r= -0.696$ ), and nitrate ( $r = -0.821$ ) in addition to chloride and TDS ( $r= -0.738$ ), moreover negative correlation was seen between

Chordate and phytoplankton ( $r = -0.705$ ). Feacel caliform affected negatively with ammonia ( $r=-0.710$ ) and TDS ( $r= -0.704$ ).

**Table 7: Correlation matrix of Mahmoudia canal water parameters.**

Correlations																	
	Turbidity	pH	EC	TDS	TSS	NH <sub>3</sub>	COD	BOD	Hardness	T. Alkalinity	Cl	PO <sub>4</sub>	NO <sub>3</sub>	NO <sub>2</sub>	Si	SO <sub>4</sub>	Temp.
Turbidity	1																
pH	.542	1															
EC	-.539-	-.609-	1														
TDS	.673*	.338	-.668.*	1													
TSS	.486	-.229-	.218	.208	1												
NH <sub>3</sub>	.291	.034	-.514-	.694*	.137	1											
COD	-.520-	-.684.*	.578	-.434-	.209	-.394	1										
BOD	.248	-.024	.128	.297	.231	-.141	.21	1									
Hardness	-.501-	-.458-	.884**	-.583-	.211	-.34	.27	-.13	1								
T. Alkalinity	-.408-	-.492-	.747**	-.169-	.074	-.12	.26	.21	.78**	1							
Cl	-.357-	-.344-	.863**	-.738.*	.349	-.575	.43	-.006	.786**	.383	1						
PO <sub>4</sub>	.852**	.647*	-.662.*	.631	.170	.392	-.48	.258	-.696.*	-.514	-.534-	1					
NO <sub>3</sub>	-.702*	-.726*	.897**	-.584	.019	-.457	.578	.152	.828**	.822**	.628	-.821**	1				
NO <sub>2</sub>	-.682.*	-.406-	.780**	-.420-	-.21	-.32	.41	.282	.709*	.852**	.484	-.60	.842**	1			
Si	.181	-.349	.367	.084	.458	-.01	.52	.660*	.017	.269	.196	.274	.199	.253	1		
SO <sub>4</sub>	-.462	-.494	.889**	-.662*	.320	-.421	.60	.25	.741*	.521	.869**	-.469	.728*	.675*	.477	1	
Temp.	-.100-	-.696.*	.446	-.340-	.396	-.350-	.425	.134	.257	.107	.440	-.418-	.460	-.013-	.235	.336	1

\*. Correlation is significant at the 0.05 level (2-tailed). \*\* Correlation is significant at the 0.01 level (2-tailed).

## CONCLUSION

Monitoring the fresh water regularly can help to identify the sources and fates of contaminants. This would mitigate outbreak of health disorders and the detrimental impacts on the aquatic ecosystem. The present paper presented comprehensive water quality information of Mahmoudia canal and analyzed sixteen physicochemical parameters at the eight sampling stations during the year 2018 using multivariate statistical techniques. The concentrations of some heavy metals were investigated at most sites of the water were free from toxic metals and below the detection limits except Aluminum ions exceeded the background level with values 0.395 and 0.355 mg/L, in both town of Kafr El Dawar and town of Abou Hommos sites. The concentration of pesticides was undetectable indicating that the canal is free from them and hence safe for drinking purposes. Water Quality Index (WQI) was calculated for all sampling sites in four seasons. A correlation matrix assessment was carried out to check the significant relationship among biological and physicochemical parameters.

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## ARABIC SUMMARY

### تقييم جودة المياه والهائمات في ترعة المحمودية في الشمال الغربي في مصر

محمد ممدوح الفقى ، أحمد عيد البرل\* ، أحمد مبروك حنيش ، حمدى ابوطالب ، محسن يوسف عمر

تقع ترعة المحمودية في الشمال الغربي لجمهورية مصر العربية، حيث تعتبر ترعة المحمودية هي المصدر الرئيسي لمياه الشرب في محافظة الاسكندرية واجزاء من محافظة البحيرة وهي مجرى ملاحى نهري بطول ٧٧ كيلو متر، حيث تستمد مياهها من مياه نهر النيل فرع رشيد أحد فرعى نهر النيل. وفي هذه الدراسة تم رصد بعض العوامل البيئية ( الفيزيائية والكيميائية والبيولوجية ) مثل درجة الحرارة، ودرجة الشفافية والعكارة ، ودرجة الأس الهيدروجيني، وكذلك درجة التوصيل الكهربى، وتركيز الأوكسجين الذائب فى المياه، وتركيز الاكسجين المستهلك بيولوجيا وكيميائيا، وتركيز الأمونيا، وتركيز الأملاح الكلية الذائبة، وتركيز المواد العالقة، و النيتريت والنترات والنيتروجين والفوسفات وأخيرا بعض العناصر الثقيلة والمبيدات. كما تم تقدير كل من المحصول القائم لكل من الهائمات النباتية والحيوانية فى عشر محطات بطول ترعة المحمودية خلال مواسم عام ٢٠١٨ م، تبدأ من فصل الشتاء فى شهر يناير وتنتهى بفصل الخريف فى شهر سبتمبر. وكان الهدف من هذه الدراسة هو الوقوف على التغييرات فى نوعية المياه وكذلك دراسة العلاقات بين المتغيرات الفيزيائية والكيميائية والبيولوجية فى البيئة المائية لمنطقة الدراسة. وقد تمت معالجة النتائج إحصائيا باستخدام معامل التنوع لتقدير المحصول القائم لمجمعات الهائمات الحيوانية، كما تم معالجة النتائج الفيزيائية والكيميائية والبيولوجية أيضا باستخدام معامل سيرمان، لبيان العناصر المؤثرة باستخدام برنامج SPSS 22.0. كما أظهرت النتائج أن :- متوسط الدليل النوعى للمياه يتراوح بين ٧٥,٠١ (فقيرة فى جودة المياه) الى ١٣٢,٥٧ (فقيرة جدا فى جودة المياه). وقد سجلت الهائمات النباتية ثلاث مجموعات هي الطحالب الخضراء المزرققة والطحالب الخضراء والدياتومات، حيث ان اعلى كثافة سنوية سجلت فى مدينة ابو حمص بينما اقل كثافة سجلت فى قرية زاوية غزال. حيث سادت مجموعة الدياتومات على المجموعتين الاخرتين. كما ان اعلى كثافة لها سجلت فى فصل الشتاء ٣٤٦٠,٢ وحده / ملى و اقل كثافة كانت فى فصل الصيف ٨٣١,٦ وحده / ملى. سجلت الهائمات الحيوانية ستة مجموعات هي الاوليات الحيوانية ومجذافية الارجل والعجليات الدوارة و القشريات الصدفية والديدان الخيطية والحبيبات. كما ان اعلى كثافة لها سجلت فى فصل الصيف ٦٠٠٠ كائن / متر مكعب و اقل كثافة كانت فى فصل الشتاء ٣٢٠٠ كائن / متر مكعب.